



HAL
open science

Effects of tritiated water on locomotion of zebrafish larvae: a new insight in tritium toxic effects on a vertebrate model species

Caroline Arcanjo, Christelle Adam-Guillermin, Sophia Murat El Houdigui, Giovanna Loro, Claire Della-Vedova, Isabelle Cavalie, Virginie Camilleri, Magali Floriani, Beatrice Gagnaire

► To cite this version:

Caroline Arcanjo, Christelle Adam-Guillermin, Sophia Murat El Houdigui, Giovanna Loro, Claire Della-Vedova, et al.. Effects of tritiated water on locomotion of zebrafish larvae: a new insight in tritium toxic effects on a vertebrate model species. *Aquatic Toxicology*, 2020, 219, pp.105384. 10.1016/j.aquatox.2019.105384 . hal-02635504

HAL Id: hal-02635504

<https://hal.science/hal-02635504>

Submitted on 27 May 2020

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial - NoDerivatives 4.0 International License

1 **Effects of tritiated water on locomotion of zebrafish larvae: a new insight in tritium toxic**
2 **effects on a vertebrate model species**

3 Caroline Arcanjo¹, Christelle Adam-Guillermin², Sophia Murat El Houdigui¹, Giovanna Loro¹,
4 Claire Della-Vedova¹, Isabelle Cavalie¹, Virginie Camilleri¹, Magali Floriani¹, Béatrice
5 Gagnaire¹

6

7 ¹Institut de Radioprotection et de Sûreté Nucléaire (IRSN), PSE-ENV/SRTE/LECO, Cadarache,
8 13115, Saint-Paul-lez-Durance, France

9 ²Institut de Radioprotection et de Sûreté Nucléaire (IRSN), PSE-SANTE\SDOS\LMDN,
10 Cadarache, 13115, Saint-Paul-lez-Durance, France

11

12 **Corresponding authors:**

13 arcanjo.caroline@hotmail.fr and beatrice.gagnaire@irsn.fr

14

15 **Abstract**

16 Tritium (³H), a radioactive isotope of hydrogen, is ubiquitously present in the
17 environment. In a previous study, we highlighted a mis-regulation of genes involved in
18 muscle contraction, eye transparency and response to DNA damages after exposure of
19 zebrafish embryo-larvae from 3 hpf to 96 hpf at 0.4 and 4 mGy/h of tritiated water (HTO).
20 The present study aimed to link this gene mis-regulation to responses observed at higher
21 biological levels. Zebrafish embryo-larvae were exposed to 0.4 and 4 mGy/h. Analyses on
22 spontaneous tail movement, locomotor activity and heart rate were performed. Histological
23 sections of eyes were made to evaluate the impact of HTO on eye transparency and whole
24 embryo immunostainings were realized to assess DNA double strand breaks repair using
25 gamma-H2AX foci. We found a decrease of basal velocity as well as a decrease of response in
26 96 hpf larvae exposed at 0.4 mGy/h after a tactile stimulus as compared to controls.
27 Histological sections of larvae eyes performed after the exposure to 4 mGy/h did not show
28 obvious differences in lens transparency or retinal development between contaminated and

29 control organisms. Gamma-H2AX foci detection revealed no differences in the number of
30 gamma-H2AX foci between contaminated organisms and controls, for both dose rates.
31 Overall, results highlighted more detrimental effects of HTO exposure on locomotor
32 behavior in 96 hpf larvae exposed at the lowest dose rate. Those results could be linked to
33 mis-regulation of gene involved in muscle contraction found in a previous study at the same
34 dose rate. It appears that not all effects found at the molecular scale were confirmed using
35 higher biological scales. These results could be due to a delay between gene expression
36 modulation and the onset of physiological disruption or homeostatic mechanisms to deal
37 with tritium effects. However, crossing data from different scales highlighted new pathways
38 to explore, i.e. neurotoxic pathways, for better understanding HTO effects on organisms.

39

40 **Keywords**

41 Tritiated water; Zebrafish embryo-larvae; locomotor-related behavior ; heart-beats ; eye
42 histology; DNA double strand breaks

43

44 **1. Introduction**

45 Organisms are largely exposed to radionuclides originating from natural or anthropic
46 sources. Among them, tritium (^3H) can be found in the environment coming from both
47 sources. As a radioactive isotope of hydrogen, tritium enters in the water cycle and
48 therefore is present in all environmental compartments (Adam-Guillermin et al., 2010). The
49 decay of tritium, following a half-time of 12.32 years, emits a beta particle of low energy
50 (mean = 5.7 keV) and low penetration range (mean = 0.56 μm in water) (HPA, 2007). This last
51 property implies that the energy deposition in tissues is very localized and gives to tritium a
52 high ionization capacity that can be considered negligible for external exposure (Mathur-De
53 Vre and Binet, 1984). However, if tritium is integrated in organism tissues, the effects could
54 be more detrimental than external exposure.

55 The tritium activity background, i.e. without nuclear facilities influences, in southern
56 France rivers ranges from 0.12 to 0.86 Bq/L (Ducros et al., 2018). In marine environment, the
57 background of tritium activities ranges from 0.02 to 0.15 Bq/L (Eyrolle-Boyer, 2017). In

58 France, tritium is mainly discharged in the environment by nuclear power plants and nuclear
59 reprocessing plants in normal operations in the form of tritiated water (HTO) or tritiated
60 hydrogen (HT). The average activity measured in nuclear power plants liquid effluents in
61 normal operations is of 2×10^6 Bq/L making tritium one of the most discharged radionuclides
62 in terms of activity compared to other radionuclides (Adam-Guillermin et al., 2010).
63 However due to the tritium properties of rapid dilution in the water cycle, activities
64 measured in rivers downstream of the nuclear facilities range from 1 to 65 Bq/L (Eyrolle-
65 Boyer, 2017). In the English Channel, downstream of the nuclear facilities tritium activity is
66 about 10 Bq/L and gradually decrease along the coast (Eyrolle-Boyer, 2017). In the future,
67 the development of new nuclear facilities, such as European Pressurized Reactor (EPR) or
68 International Thermonuclear Experimental Reactor (ITER), is expected to increase tritium
69 discharge in environment. That is why, understanding the risks of tritium on aquatic
70 organisms is of growing concern in the scientific community.

71 The effects of tritiated water exposure were studied over a large range of species, from
72 mammals through the effects on reproduction in mice or rats (Cahill and Yuile, 1970; Laskey
73 et al., 1973; Lowry Dobson and Cooper, 1974) and human cells to assess DNA damages
74 (Gonen et al., 2014) to aquatic invertebrates or vertebrates (Adam-Guillermin et al., 2012).
75 For those organisms, studies ranged from *in vitro* (Stuart et al., 2016) to *in situ* (Gagnaire et
76 al., 2017) through laboratory studies mainly focusing on developmental and reproductive
77 effects (Adam-Guillermin et al., 2012). Despite the wide range of species in which effects of
78 HTO exposure were studied, there is still a lack of knowledge of tritium modes of action at
79 the molecular scale, the transgenerational effects, the adaptation or the behavior of
80 organisms (Adam-Guillermin et al., 2012, 2010).

81 The present work is part of a project which aims to identify the modes of action of
82 tritiated water. Explaining that activities chosen for exposures are widely above those found
83 in natural environment, under nuclear facilities influence or not. To reach this objective, an
84 aquatic vertebrate model was used, the zebrafish, *Danio rerio*. The embryo larval stages
85 were chosen, as early developmental stages are supposed to be more sensitive to pollutant
86 and as the effects of pollutants on this early stage could induce detrimental effects during
87 the rest of the fish development (Frayssé et al., 2006). Furthermore, these early stages are
88 recommended in OECD guidelines for the testing of chemicals (OECD, 2013). Zebrafish is a

89 very convenient laboratory model especially when the use of the embryo-larvae stages is
90 required, as in this project. Indeed, adults can spawn hundreds of transparent eggs in a
91 single spawning (Spence et al., 2008). The genome of this species is fully sequenced, and a
92 wide variety of mutants are now available (Scholz et al., 2008), which is useful when focusing
93 on modes of action. Furthermore, zebrafish is an emerging model for studying behavior and
94 many tools are now available to assess plenty of behavioral parameters in adult and embryo-
95 larvae of zebrafish (Ahmad et al., 2012). In this project, a previous study was conducted and
96 focused on the effects of HTO exposure on zebrafish early life stages at the molecular scale
97 using a transcriptomic analysis by mRNAseq (Arcanjo et al., 2018). This study revealed an
98 effect of HTO exposure on the expression of genes involved in muscle contraction, eye
99 transparency, circadian clock, response to reactive oxygen species and repair of DNA
100 damages. The aim of the present study was to complete the results of this previous work by
101 investigating if the effects observed at the molecular scale were confirmed at higher
102 biological scales including tissue and individual, more relevant from an ecological point of
103 view. To reach this objective, 3 hpf (hours post-fertilization) zebrafish eggs were exposed to
104 the same dose rates that were previously used (i.e. 0.4 and 4 mGy/h). Both of these dose
105 rates corresponded to a band of dose rates within which there is some chance of observing
106 some deleterious effects on young fish (e.g. reduction in resistance to infection and
107 reproductive success) (ICRP, 2012). In term of activity concentrations, these two doses rates
108 corresponded to tritiated water activities of 1.22×10^5 and 1.22×10^6 Bq/mL, respectively. As
109 the first part of the study highlighted differential expression of genes involved in muscle
110 contraction (skeletal and cardiac) (Arcanjo et al., 2018), 24 hpf embryos to 96 hpf larvae
111 were challenged with behavioral analyses including spontaneous tail movements, heart rate
112 and free-swimming behavior. Histological observations of eyes from 48 hpf to 96 hpf were
113 also performed to make the link with modulation of genes involved in eye transparency.
114 Finally, a whole mount immunostaining was performed to detect gamma-H2AX foci, an early
115 marker of double strand break repair process (Rogakou et al., 1998), as some genes involved
116 in DNA repair pathways were found mis-regulated in a previous study (Arcanjo et al., 2018).

117

118 **2. Materials and methods**

119 ***2.1. Fish maintenance and eggs production***

120 Adult wild type zebrafish (AB genetic background) were maintained in a Tecniplast©
121 ZebTEC system in 8L tanks (20 adults per tank) with a temperature of 28°C, males and
122 females being kept separated. A photoperiod of 12/12h dark-light cycles was used. Eggs
123 were produced by mating couples in 1.7L Breeding Tank Beach Style Design (Tecniplast©) in
124 a thermoregulated room at 28°C. Males and females were kept apart in the same breeding
125 tank during the night before the mating. The morning of mating, couples were allowed to
126 mate, and eggs were sampled 30 min after the beginning of the mating in order to obtain
127 synchronised eggs. All eggs were pooled together, sorted, counted and kept in petri dish
128 with embryo medium (60mg/L of Instant Ocean© sea salt diluted in ultrapure water
129 (Westerfield, 1995)), at $28 \pm 0.5^\circ\text{C}$ until the beginning of the experiments.

130

131 ***2.2. Eggs contamination and sampling***

132 Tritiated water source (97% purity, 185 MBq, PerkinElmer, Courtaboeuf, France) was diluted
133 in embryo medium to obtain final activity concentrations of 1.22×10^5 and 1.22×10^6 Bq/mL.
134 At 3 hpf, eggs were contaminated respecting a density of 5 eggs per 2 mL of HTO solution in
135 three experimental units (three replicates). To perform the behavioural and physiological
136 experiments, the histological sections and the whole mount immunostaining, 1620 eggs
137 were used for both conditions, control and HTO-contaminated (135 eggs * 3 replicates * 2
138 treatments * 2 dose rates). The eggs were kept in the dark at $27.7 \pm 0.3^\circ\text{C}$ in a thermo-
139 regulated incubator confined inside a glove box. Control eggs were placed in embryo
140 medium and kept in the same manner as contaminated organisms ($27.8 \pm 0.07^\circ\text{C}$) in the
141 same room outside the glove box. Water in contaminated and control experimental units
142 was renewed once (at 72h) during the experiment to ensure water quality and constant
143 tritiated water activity. Samples of contamination medium were made to monitor tritium
144 activity along the experiments in the three experimental units. These samples were analysed
145 by liquid scintillation counting. Briefly, samples were diluted with tritium-free water (Les
146 Abatilles, Arcachon, France) to obtain a volume of 1 mL. A liquid scintillation cocktail (Ultima
147 Gold LLT, PerkinElmer) was then added to a final volume of 20 mL. Samples were analysed
148 using a Quantulus 1220 (DL = 5 Bq/L, PerkinElmer, WinQ software) and spectrum were
149 analysed using the EASY View Spectrum analysis Program (PerkinElmer). To avoid cross-
150 contamination between the two HTO activity concentrations, experiments were conducted

151 separately. The statistical analyses performed to compare mean activity concentration
152 between experimental units showed no differences for both activities tested in this study.
153 Measured activity concentrations along experiments and for all experimental units were of
154 $1.10 \times 10^5 (\pm 5.72 \times 10^3)$ Bq/mL and $1.35 \times 10^6 (\pm 8.89 \times 10^4)$ Bq/mL respectively for the lowest and
155 the highest activity concentration tested.

156 Concerning the sampling of organisms needed for the behavioral and histological
157 analyses as well as for the gamma-H2AX immunostaining, contaminated organisms were
158 washed three times in non-tritiated embryo medium to eliminate tritium contamination
159 around the eggs that could lead to over-estimate radiological dose. Control organisms were
160 handled in the same way. This protocol was optimised for zebrafish (Arcanjo et al., 2019).

161

162 **2.3. Behavioral analyses**

163 All behavioral analyses were preceded by an acclimation period of 30 min for all
164 organisms (control and contaminated) in the behavioral room. Acclimation was performed in
165 a dark thermoregulated incubator (AquaLytic, Germany) at $28 \pm 1^\circ\text{C}$. Videos for the
166 assessment of spontaneous tail movements of chorionated embryos at 24 hpf and heart rate
167 of 48 hpf and 72 hpf larvae were performed under a stereomicroscope (Nikon SMZ800)
168 connected to a high-resolution camera (acA1300-60gm, Basler, Germany) under the
169 behavioural room light and temperature (28°C). Recording were performed using Media
170 Recorder Software (v.4.0.542.1, Noldus Information Technology, Netherlands) and the
171 analysis of videos was performed using the DanioScope Software (v.1.1.110, Noldus
172 Information Technology, Netherlands). Concerning the swimming behavior of 96 hpf larvae,
173 the DanioVision Observation Chamber (Noldus Information Technology, Netherlands)
174 coupled with the DanioVision Temperature Control Unit (Noldus Information Technology,
175 Netherlands) at $28^\circ\text{C} (\pm 0.1)$ were used and videos were analysed using the EthoVision XT
176 software (v.13.0.1216, Noldus Information Technology, Netherlands). All videos used for the
177 swimming behavior analyses were performed in the dark under infra-red light. Results for all
178 behavioral tests were compared to their respective control.

179

180 *2.4.1. Spontaneous tail movements of 24 hpf embryos*

181 For the assessment of the spontaneous tail movement in 24 hpf embryos, three
182 replicates of 30 organisms were used. As the 30 organisms cannot be recorded together
183 under the stereomicroscope, replicates were divided in three sub-replicates of 10 organisms.
184 After the acclimation period, sub-replicates were placed on a petri-dish containing 1%
185 agarose gel (Sigma-Aldrich), then 20 μ L of methyl cellulose was added to avoid movement of
186 eggs during the recording time. Organisms were then placed under the stereomicroscope
187 and after an acclimation period of 2 min before the videos were recorded (5 min of video,
188 frame rate = 25 Hz). Control and contaminated replicates were alternated. All the recordings
189 of the spontaneous movement in 24 hpf embryos took place between 10:20 am and 01:00
190 pm. For the statistical analysis, the burst activity was used. It corresponds to the percentage
191 of time the embryo was moving along the duration time of the video. Based on the data
192 distribution, values below 1 and above 40 were considered as outliers and were excluded
193 from the data set.

194

195 *2.4.2. Heart rate of 48 hpf and 72 hpf larvae*

196 For the heart rate analysis, protocol was almost the same as described for the activity
197 assessment, except that three replicates of 10 organisms were used for each stage. These
198 replicates were divided in three sub-replicates of 5 organisms. After the acclimation period,
199 sub-replicates were placed on a petri-dish containing 1% agarose gel (Sigma-Aldrich), then
200 20 μ L of methyl cellulose and 1 μ L of MS-222 final concentration 0.02 % (w.v.) (Tricaine
201 methanesulfonate, Molekula) were added to avoid movement of larvae during the recording
202 time. After the end of video, larvae were placed in clean embryo medium and started to
203 swim normally. When anesthetised, organisms were then placed under the
204 stereomicroscope and, after an acclimation period of 2 min, the videos were recorded (30
205 sec of video, frame rate = 25 Hz). Control and contaminated replicates were alternated. All
206 the recordings of the heart rate in 48 hpf and 72 hpf larvae took place between 10:15 am
207 and 11:30 am. For the statistical analysis, based on the data distribution, a number of beats
208 per second below 1.5 and above 3.5 were considered as outliers and were excluded from the
209 data set.

210

211 *2.4.3. Swimming behavior of 96 hpf larvae*

212 The swimming behavior of the 96 hpf larvae was assessed using 48 organisms per
213 treatment distributed in four 24-wells plates. Organisms were placed individually in each
214 well with 2 mL of embryo medium; controls and contaminated animals were placed in the
215 same plate. All four plates were prepared few minutes before the beginning of their
216 respective recording time. Once placed in the DanioVision Observation Chamber, an
217 acclimation period of 10 min was respected before the beginning of the recording. Videos of
218 10 min were recorded with a tapping stimulus (level=8, on the side of the well plate) at t+5
219 min after the beginning of the recording. All the recordings of the swimming behavior in 96
220 hpf larvae took place between 11:00 am and 12:40 pm. Analysis were made between control
221 and contaminated during the first five minutes of the video to assess the swimming activity
222 basal level. The response to the tactile stimulus was assessed by comparing responses 1
223 second before and 1 second after the stimulus for all treatment. Then the delta of the
224 response (after-before) was used to compare the response to the tactile stimulus between
225 control and contaminated larvae.

226

227 ***2.5. Histological analysis of eyes***

228 The histological sections of zebrafish eyes from 48 hpf to 96 hpf were performed
229 after the exposure to 4 mGy/h of HTO only. Sections of 400 nm were realized using an
230 ultramicrotome UCT (Leica Microsystems GmbH, Wetzlar, Germany) and stained with
231 toluidine blue. Sections were then observed with an optic microscope (DM750, Leica) with
232 ICC50 camera (Leica) coupled with the LAS EZ software. Images focusing on retina and lens
233 were made using the 100x objective. Both eyes for two larvae were used for the three stages
234 studied representing 12 organisms.

235

236 ***2.6. Whole mount gamma-H2AX immunostaining***

237 *2.6.1. Immunostaining*

238 The whole mount gamma-H2AX immunostaining were performed using at least 20
239 embryos at 24 hpf for each treatment (control and contaminated). After the sampling, eggs
240 were manually dechorionated and fixed in 2 mL of PFA 4% in PBS 1X (Sigma-Aldrich) during 3
241 h at 4°C. Once fixed, embryos were kept in 100% methanol (VWR International) at -20°C until
242 the beginning of the staining.

243 The staining begins by the rehydration of embryos with successive washes with
244 decreasing concentrations of methanol in PBS 1X (70%; 50%; 25% methanol). Embryos were
245 then permeabilized with Proteinase K in PTW buffer (10 µg/mL final concentration) for 3 min
246 at room temperature. After a brief wash in PTW buffer, they were fixed in PFA 4% in PBS 1X
247 (Sigma-Aldrich) for 20 min at 4°C. The samples were blocked in a 0.2% BSA and 1% DMSO
248 PBST solution for 3 h at room temperature, and then incubated with the primary antibody
249 (1:200 gamma-H2AX in blocking solution, GeneTex GTX127342) at 4°C overnight. The
250 samples were washed five times in PTW buffer, incubated with the secondary antibody
251 (1:500 Sigma-Aldrich anti rabbit-FITC F0382) 2 h at room temperature and then accurately
252 washed with PTW buffer. The samples were kept protected from light at 4°C until analysis by
253 confocal microscopy.

254

255 *2.6.2. Confocal observations*

256 Confocal microscopy analyses were performed using a laser scanning confocal
257 microscope (Carl Zeiss, France) using a 20x dry objective (N.A. 0.8), with a pinhole setting of
258 1 A.U. For FITC fluorescence, excitation was performed at 488 nm and emission acquired
259 between 510/550 nm. For DAPI detection, samples were stained in a 0.1% Tween and
260 300nM DAPI PBS solution for 1 h and 30 min, excitation was performed at 405 nm and
261 detection between 420 and 500 nm. The samples were mounted on 35 mm imaging dish
262 with a polymer coverslip bottom (Ibidi) and covered by a 2% methyl-cellulose solution.

263

264 *2.6.3. Image analysis*

265 The images acquired at the confocal microscope were analysed using the open-
266 source software Fiji ImageJ. For each image, the smooth filter was applied (this filter
267 replaces each pixel with the average of its 3x3 neighbourhood) to increase the ratio

268 signal/noise, and then the level of background was chosen (values between 5 and 10).
269 Several regions of interest (ROI) were designed on the nuclei detected in the DAPI channel,
270 using the ImageJ tool ROI manager. The ROI were transferred and combined with the FITC
271 channel. The function Find Maxima was used to identify and count the foci; this algorithm
272 determines local maxima that pass the threshold (noise tolerance) in the selected area. The
273 values derived from the initial background analysis were used as a threshold. The ratio
274 between the number of maxima identified by the function and the number of selected nuclei
275 was reported as foci per nucleus.

276

277 **2.7. Statistical analyses**

278 All statistical analyses were performed using the R software (v 3.4.3) (R Core Team,
279 2017). Behavioral parameters, i.e. tail spontaneous movements, heart rate and locomotor
280 activity, were analysed using linear mixed effect models fitted with the *nlme* package
281 (Pinheiro et al., 2017). Treatment and time (at which the video was recorded reported as
282 “time effects” in the text) were included as fixed effects while replicates were included as
283 random effects. Residuals normality and homogeneity assumptions were assessed
284 employing QQplot and fitted values vs. standardized residuals plot. Since they were not
285 satisfied random permutation tests were finally applied using the *pgirmess* package
286 (Giraudoux et al., 2018).

287 HTO activity in water and gamma-H2AX foci were analysed using Student t tests, or
288 Wilcoxon tests if validity condition were not satisfied, and using two ways ANOVA. Normality
289 and homogeneity assumptions were assessed using Shapiro-Wilk tests and Bartlett tests
290 respectively. Hypotheses were bilateral, and significance level was set to 0.05.

291

292 **3. Results**

293 **3.1. Behavioral and physiological analysis**

294 In 24 hpf embryo, the spontaneous movements of the tail were not significantly
295 different in controls and in contaminated organisms for both dose rates (Figure 1). A

296 significant effect of the time was found for both dose rates and treatment as a diminution of
297 tail movement (data not shown).

298 The heart rate of 48 hpf and 72 hpf larvae were also analysed. Beats per second (BPS)
299 in control and contaminated organisms were not significantly different for both stages and
300 dose rates (Figure 2). A significant effect of the time was highlighted for both dose rates in
301 48 hpf larvae and only for 4 mGy/h in 72 hpf larvae as an increase of BPS (data not shown).

302 Concerning the swimming behavior of 96 hpf larvae after a 0.4 mGy/h exposure,
303 contaminated organisms showed a significant lower velocity (mm/s) during the first five
304 minutes of the tracking videos than control organisms (Figure 3A). This difference was not
305 found in 96 hpf larvae exposed to 4 mGy/h (Figure 3B). A second experiment was conducted
306 with 96 hpf larvae exposed at 0.4 mGy/h and the same trend was highlighted (Figure S1).
307 Following the tapping stimulus (i.e. one second before and one second after), control and
308 contaminated 96 hpf larvae showed significant differences in velocity (mm/s). Indeed, the
309 velocity of 96 hpf larvae was significantly lower one second before the tapping stimulus than
310 one second after in control and contaminated organisms for both dose rates (Figure 4). The
311 delta of the response to the stimulus (i.e. after-before) between control and contaminated
312 organisms was compared for both dose rates. Results showed that after an exposure to 0.4
313 mGy/h, the delta of the response in contaminated 96 hpf larvae was significantly lower than
314 the delta observed in control organisms (Figure 5). After an exposure to 4 mGy/h, no such
315 difference was found (Figure 5).

316

317 ***3.2. Histological observations of eyes***

318 The histological observations were performed on larvae eyes from 48 hpf to 96 hpf
319 exposed to a dose rate of 4 mGy/h. The comparison between pictures did not show notable
320 alterations in the retinal structure when comparing control and contaminated organisms for
321 all stages (Figure S2). Indeed, as expected when normal development occurs, the five layers
322 composing the retina were not distinguishable from each other in 48 hpf larvae but well
323 defined in 72 hpf and 96 hpf larvae (Figure S2). When focusing on the lens, no remarkable
324 difference appeared in lens development between 4 mGy/h exposed larvae and controls for
325 all the three stages. In 48 hpf larvae, lens was fully detached from the ectoderm and showed

326 a mono layer of epithelial cells on the anterior part of the lens while in the central part
327 elongated primary lens fibre were visible as a concentric layer (Figure 6). On later stages (i.e.
328 72 hpf and 96 hpf larvae), the lens structure became more homogenous, with cells forming a
329 concentric structure still visible in 72 hpf larvae but not in 96 hpf larvae (Figure 6).

330

331 **3.3. Whole mount gamma-H2AX immunostaining**

332 For both dose rates, the comparison between control and contaminated organisms
333 was performed using the foci per nucleus ratio. For control and samples exposed to 0.4
334 mGy/h, a total of 850 and 779 cells were counted, respectively. A similar number of cells
335 were counted in control and 24 hpf embryos exposed to 4 mGy/h (824 for control and 936
336 for contaminated samples). For both dose rates, no significant difference was highlighted
337 between control and contaminated 24 hpf embryos (Figure 7).

338

339 **4. Discussion**

340 **4.1. Behavior and heart rate**

341 *4.1.1. HTO exposure did not change spontaneous tail movement*

342 At 24 hpf, the spontaneous movement of the zebrafish embryo tail was investigated.
343 Results did not show significant differences between control and contaminated organisms
344 for both dose rates. These results suggested that genes coding for proteins involved in
345 muscle contraction found, via a transcriptomic analysis, to be affected in 24 hpf eggs for
346 both dose rates (Arcanjo et al., 2018) did not induce any effect on spontaneous tail
347 movement at the same developmental stage. The spontaneous movements of the tail
348 originate from the spinal cord and are dependant of functional motoneuron innervation
349 (Brustein et al., 2003) as well as muscle contraction. Furthermore, a well-known
350 neurotoxicant such as chlorpyrifos was shown to increase tail movements in zebrafish
351 embryo from 24 to 26 hpf from 0.625 to 10 mg/L (Selderslaghs et al., 2010). Our result
352 suggests that this pathway did not seem to be affected by HTO exposure in 24 hpf eggs for
353 both dose rates tested, questioning the neurotoxic effect of HTO. However, genes coding for
354 proteins involved in neurotransmitter and Ca²⁺ transport were modulated in 24 hpf eggs

355 after exposure to 0.4 and 4 mGy/h of HTO (Arcanjo et al., 2018). Furthermore, in 24 hpf eggs
356 exposed to 4 mGy/h of HTO an over-expression of the *ache* gene, coding for the
357 acetylcholinesterase protein that play a role in muscle relaxation or muscular and neuronal
358 development (Behra et al., 2002), was found compared to control (Gagnaire et al, in prep).
359 Together, these results suggest that the neuro-muscular functions could be impaired after
360 HTO exposure and it would be interesting to perform further investigations with older
361 developmental stages. Indeed, as genes mis-regulation were found at 24 hpf, it was probably
362 too soon to see any phenotypical impairments at the same time. That is why we also
363 performed locomotor analysis on 96 hpf larvae to assess the latency between the gene
364 expression impairments and the phenotypical observations.

365 Interestingly, the statistical analysis performed on the spontaneous tail movement
366 highlighted a significant effect of time for both HTO exposures; meaning that the time during
367 which experiments were made has a significant effect on the measured parameter (burst
368 activity). Indeed, the first spontaneous movement appeared around 17 hpf. In
369 dechorionated embryos, the spontaneous movements reach a peak around 19 hpf and then
370 slowly decrease by 26 hpf (Saint-Amant, 2006). Such a decrease of the spontaneous tail
371 movement was also reported in literature in control organisms (Huang et al., 2010; Wang et
372 al., 2013) confirming that time influence this parameter. In the present study, as
373 experiments were performed on stages between 24 hpf and 27 hpf, it is not surprising to
374 find a significant effect of time in the obtained data.

375

376 *4.1.2. HTO exposure did not change the number of beats per seconds*

377 After an exposure to 0.4 mGy/h of HTO, no significant differences in BPS were
378 highlighted between control and exposed organisms for both stages. The same result was
379 found for the highest dose rate of 4 mGy/h. Furthermore, the statistical analyses showed a
380 significant effect of time for both dose rate when 48 hpf larvae were investigated and only
381 for the highest dose rate for the 72 hpf larvae suggesting that the time at which the
382 experiments are recorded could influence the heart rate. Together, the results found for the
383 cardiac activity after HTO exposure suggested no effect of HTO on this parameter for the
384 developmental stages investigated. As previously said, it appears that the down-regulation

385 of genes involved in both muscle contraction and event leading to the muscle contraction
386 found in a previous study in 24 hpf eggs (Arcanjo et al., 2018) had no effects on cardiac
387 activity on 48 hpf and 72 hpf larvae. These results could indicate that the organisms can cope
388 with tritium toxicity with early gene expression changes, indicating the low toxicity of
389 tritium, at least for the considered life stages and exposure period. Using gamma radiation, it
390 is possible to observe effects on heart rate with dose as low as 0.5 mGy/h in 48 hpf larvae
391 (Murat El Houdigui et al, under review). To go further, it would be interesting to perform
392 dedicated experiments to find the moments of the day were larvae show constant BPS to
393 avoid time effect.

394

395 *4.1.3. HTO exposure induces effects on locomotor behavior*

396 To get insight into the swimming abilities of 96 hpf larvae after HTO exposure, a free-
397 swimming behavioral test was performed in 96 hpf larvae, followed by a tactile stimulus on
398 the side of the plate containing organisms. Results highlighted that during the first five
399 minutes of the video tracking, the 0.4 mGy/h exposed larvae showed a mean velocity
400 significantly lower than the velocity of control larvae. The response to a tactile stimulus was
401 then investigated and showed that controls and HTO-contaminated organisms (for both
402 dose rates) responded to the tactile stimulus by a significant enhancement of their velocity
403 one second after the stimulus. The delta of the response to the tactile stimulus was also
404 studied. It appeared that the 0.4 mGy/h exposed 96 hpf larvae showed a delta of the
405 response significantly lower than control organisms.

406 The effects found on 96 hpf larvae locomotor activity could be explained by the
407 potential impact of HTO exposure on muscle structure and events leading to muscle
408 contraction. Indeed, muscles impairments as disorganization of sarcomeres myofilaments
409 (96 hpf) and disruption of genes involved in muscle contraction (24 hpf for both dose rate)
410 were reported in a previous study using the same HTO dose rates (Arcanjo et al., 2018). Also
411 in zebrafish, gamma exposure to 0.03 mGy/h and 24 mGy/h induced degradation of
412 myofibrils which were more deleterious for the higher dose rate tested (Gagnaire et al.,
413 2015). In 96 hpf larvae TEM muscle observation reveals myofibers and neuromuscular
414 junction alterations after gamma irradiation at 5 and 50 mGy/h (Murat El Houdigui et al,

415 under review). Furthermore, after HTO or gamma exposure, vertebral malformations have
416 been documented in medaka, *Oryzias latipes*, exposed to 35 mGy/h (HTO) and 18 mGy/h (γ)
417 during embryonic stages (Hyodo-Taguchi and Etoh, 1993) which could lead to locomotor
418 impairments. To explain the locomotor effects of HTO on zebrafish larvae, it would be also
419 interesting to focus on the lateral line, a mechanosensory system in fish and on the
420 Mauthner cells which mediated the trunk and tail contraction during the response to a
421 tactile stimulus (Eaton et al., 1977). Indeed, after uranium exposure, severe tissue
422 alterations were reported in the lateral line system in zebrafish (Faucher et al., 2012).

423 However, the impairments of the muscular system, both at the molecular and tissue
424 scales, cannot explain the observed effects alone, as the locomotor activity in 96 hpf larvae
425 exposed to 4 mGy/h was not affected by HTO exposure. As a proper locomotor activity
426 depends on both the muscular and the nervous systems (Drapeau et al., 2002), the potential
427 neurotoxic effect of HTO exposure has also to be considered. After gamma irradiation,
428 zebrafish larvae show neuronal as well as muscles impairments (Murat El Houdigui et al,
429 under review). After HTO exposure, a transcriptomic analysis highlight down-regulation of
430 genes coding for proteins involved in peripheral nervous system axonogenesis (*her4.3* and
431 *her4.4*) in 24 hpf embryos after exposure at 4 mGy/h (Arcanjo et al., 2018). In studies
432 focusing on chemicals such as nicotine (Parker and Connaughton, 2007; Svoboda et al.,
433 2002), caffeine (Chen et al., 2008) or ethanol (Carvan et al., 2004), effects on the locomotor
434 behavior of zebrafish larvae after a tactile stimulus, dose dependant decrease of response
435 was shown in contaminated organisms comparing to controls. The mode of action evoked to
436 explain these results are linked to an effect on the development and on the alteration of the
437 axonal pathfinding of the secondary spinal motoneurons as defects on the neuromuscular
438 junction (Chen et al., 2008; Svoboda et al., 2002).

439 Together, these results suggest that HTO exposure had more detrimental effect after
440 an exposure to 0.4 mGy/h than to 4 mGy/h on the locomotor activity in 96 hpf larvae. This
441 lack of concentration-dependant response could be due to the onset of more efficient
442 defence mechanisms following a threshold that is not reach for the lowest dose rate. Indeed,
443 in a previous study, we found more differentially expressed genes involved in DNA repair
444 pathways at 4 mGy/h than at 0.4 mGy/h (Arcanjo et al., 2018). However, at 0.4 mGy/h, the
445 events triggering the impairment of the locomotor activity in larvae remained to be

446 confirmed. Indeed, the velocity of the control larvae in the 0.4 mGy/h experiment was
447 higher than the velocity in control larvae for the experimentation with the highest dose rate.
448 A second experiment were performed at 0.4 mGy/h, and the same trend was highlight with
449 a mean velocity in control of the same range as for the first experiment. The difference in
450 control larvae mean velocity between dose rates could be explained by biological variation
451 between the pool of eggs used. Indeed, to avoid HTO cross-contamination, the experiments
452 for both dose rate were conducted separately in time, meaning that pools for both
453 experiments were not from the same spawn. This could lead to find a significant difference
454 in velocity where there is not. Nevertheless, the impact of HTO on the velocity of larvae
455 could cause delayed response to threats by impairing the escaping/avoidance behavior or
456 decrease success of prey capture and could therefore have a detrimental impact on fish
457 fitness.

458 In the literature, it is known that many variables could influence zebrafish larvae
459 locomotor behavior, i.e. larvae age, size of well, time of the day or experiment performed in
460 light or dark (MacPhail et al., 2009; Padilla et al., 2011). In the present study, we chose to
461 assess locomotor activity on 96 hpf larvae to be consistent with experiments on gene
462 expressions where genes involved in muscle contraction were mis-regulated in 24 hpf
463 embryos. Furthermore, in locomotor activity experiments performed in the dark, as in the
464 present study, response in 4 dpf, 5 dpf and 6 dpf larvae are similar to each other, which is
465 not the case in light experiments (Padilla et al., 2011), reinforcing the use of 96 hpf larvae in
466 our study. Time of the day when the experiments are performed could also influence larvae
467 locomotor response. Indeed, in 6 dpf larvae, the locomotor activity assessed in the dark
468 reached stability in early afternoon around 01:00 pm (MacPhail et al., 2009). Even though
469 our experiments were performed in late morning (from 11:00 am to 12:40 pm) when it
470 seems that locomotor activity is still stabilizing, we were able to find effects of HTO exposure
471 on larvae velocity.

472

473 **4.2. Histology of the eyes**

474 As cataract is one of the complications after beta radiation therapy in Human and
475 that cataract incidence was shown to increase in Chernobyl birds (Barron et al., 1970;

476 Mousseau and Møller, 2013), the potential development of cataract in HTO exposed
477 zebrafish larvae was investigated. Furthermore, as irregularities in the development of
478 zebrafish larvae eye were reported after gamma radiation in zebrafish larvae exposed only
479 via their parents, exposed only as larvae or both (Hurem et al., 2017), eye layer development
480 was assessed. Histological sections of eyes lens and retina were performed. As genes
481 involved in eye transparency were found down-regulated in 24 hpf eggs after an exposure to
482 4 mGy/h in a previous study (Arcanjo et al., 2018), all sections were made at this dose rate.
483 Results highlighted no distinguishable defects in larvae (from 48 hpf to 96 hpf) when
484 comparing control and contaminated organisms. Indeed, as expected in normal zebrafish
485 eyes development, in 48 hpf larvae the lens is fully detached from the cornea and show
486 different kind of cells, an anterior monolayer of epithelial cells and in more centrally the
487 primary lens fibre starting to elongated (Dahm et al., 2007). At 72 hpf, histological sections
488 of lens showed a more homogeneous appearance with lens fibre cell even more elongated
489 with degraded nucleus (Dahm et al., 2007). Finally, at 96 hpf, sections of the lens also
490 showed a homogenous pattern with no appearance of nuclei. At this stage the
491 differentiation of lens fibre cells is terminated, and lens is totally transparent (Goishi et al.,
492 2006). Concerning the histological sections of retina, no obvious differences could be seen
493 between control and contaminated organisms regardless of stages. Indeed, the five retinal
494 layers were well distinguishable in 72 hpf and 96 hpf larvae. These five layers are composed
495 of three nuclear layers (ONL, INL and GCL) and two plexiform layers (IPL and OPL) (Chhetri et
496 al., 2014). No remaining nuclei was found in lens, as already shown by histological section of
497 zebrafish eyes with cataract (Dhakal et al., 2015; Goishi et al., 2006). Therefore, these results
498 did not support the molecular alterations previously shown (Arcanjo et al., 2018).

499 Together, the results showed no effect of HTO on lens transparency or retinal
500 development at the analysed developmental stages. This results suggest that the transient
501 down-regulation of genes involved in eye transparency at 24 hpf (Arcanjo et al., 2018) did
502 not have any effect on eye development. Another hypothesis would be that toxico-dynamics
503 are different between molecular and tissular effects. To check this last hypothesis, it would
504 be interesting to follow eye development over the whole larvae development until the adult
505 stage. In addition, considering the observed behavioural changes, it would be an asset to
506 investigate behavioral tests dedicated to visual parameters as the optokinetic response

507 (OKR) or the optomotor response (OMR) in 96 hpf larvae. Indeed, at this stage the retinal
508 structure and the ocular muscles are functional, which leads to visible eye movements which
509 could indicate a behavioral changes, liable to occur even in the absence of any structural
510 defects (Chhetri et al., 2014). Indeed, the OKR was used to isolated zebrafish mutant with
511 visual defect and highlighted two mutants lacking obvious morphological difference from
512 controls (Brockerhoff et al., 1995).

513

514 **4.3. gamma-H2AX foci detection**

515 The effects of HTO exposure on DNA damages were already investigated using
516 a wide range of species as mollusc (Hagger et al., 2005; Jha et al., 2005), aquatic vertebrates
517 (Gagnaire et al., 2017; Suyama et al., 1981) or fish cell culture (Stuart et al., 2016). In the
518 present study, we focused on DNA double strand breaks in 24 hpf embryos using an
519 immunostaining method to assess the number of gamma-H2AX foci, an early marker of DNA
520 double strand break repair (Rogakou et al., 1998). Results showed that, comparing to their
521 respective controls, no significant differences were found for both dose rates. After an
522 exposure to tritium at much lower activities than in the present study, no difference
523 between control and contaminated organisms were neither found in fish cells and *in situ*
524 exposed fathead minnows (Gagnaire et al., 2017; Stuart et al., 2016) using gamma-H2AX foci
525 detection. At the same stage, i.e. 24 hpf, no DNA damages were found using the comet assay
526 for both dose rates (Gagnaire et al, in prep). Using fish cells (3B11 and FHMT-W1), no effects
527 of HTO exposure was found when investigating the gamma-H2AX foci intensity after
528 exposure from 0.01 to 100 Bq/mL (Stuart et al., 2016). After gamma irradiation, no
529 significant difference in the gamma-H2AX detection was found after exposure at dose rate
530 close to those in the present study (5 mGy/h) for the same stage while an increase of foci
531 was detected at 50 mGy/h (Murat El Houdigui et al, under review). By exposing directly ZF4
532 zebrafish cells to chronic gamma radiation, an increase of gamma-H2AX foci per nucleus was
533 found after 24h at 4 mGy/h while no effects was found at 0.4 mGy/h (Pereira et al., 2011).

534 Overall, the results found in the present study combined with those in studies
535 investigating gamma-H2AX foci detection after exposure to HTO or gamma radiation suggest
536 that organisms could cope with DNA damages potentially occurring after an exposure to

537 ionizing radiations. These results could be due to an enhancing of DNA repair pathways at
538 those dose rates to balance the effects of ionizing radiations. This hypothesis is consistent
539 with previous findings at the molecular scale highlighting an up-regulation of genes involved
540 in DNA repair pathways, in 24 hpf eggs after an exposure to HTO at 4 mGy/h (Arcanjo et al.,
541 2018). However, this hypothesis remains to be confirmed by further investigation on the
542 DNA repair pathways after ionizing radiation exposure.

543

544 **5. Conclusion**

545 In this study, observed effects at the molecular scale, i.e. genes involved in
546 muscle contraction, eye transparency and response to DNA damage, have not all been
547 confirmed at higher biological scales using stages from 24 hpf to 96 hpf. However it was
548 possible to confirm these early effects for locomotion. Results highlighted that the 0.4
549 mGy/h exposure could have more detrimental effects than the 4 mGy/h exposure on the
550 locomotor activity. These results could be partly explained by the impact on muscle
551 (structure and genes) previously highlighted for both dose rate, but not entirely. Therefore,
552 an impact on the nervous system coupled with muscular impairments could be more
553 relevant for the explanation of effects on the locomotor behavior. Furthermore, the ability
554 to cope with damages after HTO exposure could also be advanced to explain the observed
555 effect on locomotion. Concerning the other parameters followed in this study, it could be
556 interesting to explore the responses by using older larvae as there may be a delay between
557 gene expression changes and the onset of physiological disruption or homeostatic
558 mechanisms onset at those early stages which could be exceeded along time. Despite that,
559 the present study and the whole project related to, allowed to link responses observed at
560 different organization levels, from molecule to individual, contributed to improve the
561 knowledge of tritiated water exposure effects. From an ecological risk assessment point of
562 view, the present study did not highlight detrimental effects at the individual scale with
563 tritium activity as high as 1.22×10^9 Bq/L. Activities used in the present study are widely
564 higher than those found in aquatic environment suggesting that environmental exposure to
565 tritium could lead to poor detrimental effects in organisms. However, it is not excluded that
566 chronic exposure to environmental activities of tritium could lead, over time, to detrimental

567 effects at the individual level. Indeed, at the same activities as is the present work, molecular
568 effects were highlight and could, with time, generate detrimental effects.

569

570

571 **Acknowledgements**

572 The authors want to thank Marie-Christine THIBAUD and H el ene JAVOT from the ZoOM
573 platform (CEA, Cadarache) for the confocal microscope formation. Support for this
574 microscopy equipment was provided by the R egion Provence Alpes C ote d'Azur, the Conseil
575 General of Bouches du Rh one, the French Ministry of Research, the CNRS and the
576 Commissariat   l'Energie Atomique et aux Energies Aleternatives. This work was supported
577 by the GGP research program supported by IRSN (Institut de Radioprotection et de S uret e
578 Nucl eaire) and EDF (Electricit e de France).

579

580 **References**

581 Adam-Guillermin, C., Antonelli, C., Bailly du Bois, P., Beaugelin-Seiller, K., Boyer, P., Fi vet, B.,
582 Garnier-Laplace, J., Gurriaran, R., Le Dizes-Maurel, S., Maro, D., Masson, M., Pierrard,
583 O., Renaud, P., Roussel-Debet, S., 2010. Le tritium dans l'environnement, in: Le Livre
584 Blanc Du Tritium - Autorit e de S uret e Nucl eaire. pp. 44–110.

585 Adam-Guillermin, C., Pereira, S., Della-vedova, C., Hinton, T., Garnier-laplace, J., 2012.
586 Genotoxic and Reprotoxic Effects of Tritium and External Gamma Irradiation on Aquatic
587 Animals. *Rev. Environ. Contam. Toxicol.* 67–103. [https://doi.org/10.1007/978-1-4614-](https://doi.org/10.1007/978-1-4614-3414-6)
588 3414-6

589 Ahmad, F., Noldus, L.P.J.J., Tegelenbosch, R.A.J., Richardson, M.K., 2012. Zebrafish embryos
590 and larvae in behavioural assays. *Behaviour* 149, 1241–1281.
591 <https://doi.org/10.1163/1568539X-00003020>

592 Arcanjo, C., Armant, O., Floriani, M., Cavalie, I., Camilleri, V., Simon, O., Orjollet, D., Adam-
593 Guillermin, C., Gagnaire, B., 2018. Tritiated water exposure disrupts myofibril structure
594 and induces mis-regulation of eye opacity and DNA repair genes in zebrafish early life

595 stages. *Aquat. Toxicol.* 200, 114–126. <https://doi.org/10.1016/j.aquatox.2018.04.012>

596 Arcanjo, C., Maro, D., Camilleri, V., Cavalié, I., Simon, O., Beaugelin-seiller, K., Carasco, L.,
597 Orjollet, D., Adam-guillermin, C., 2019. Assessing tritium internalisation in zebrafish
598 early life stages : Importance of rapid isotopic exchange. *J. Environ. Radioact.* 203, 30–
599 38. <https://doi.org/10.1016/j.jenvrad.2019.02.009>

600 Barron, A., McDonald, J.E., Hughes, W.F., 1970. Long-term complications of beta-radiation
601 therapy in ophthalmology. *Trans. Am. Ophthalmol. Soc.* 68, 113–128.

602 Behra, M., Cousin, X., Bertrand, C., Vonesch, J.-L., Biellmann, D., Chatonnet, A., Strähle, U.,
603 2002. Acetylcholinesterase is required for neuronal and muscular development in the
604 zebrafish embryo. *Nat. Neurosci.* 5, 111.

605 Brockerhoff, S.E., Hurley, J.B., Janssen-bienholdt, U., Neuhauss, S.C.F., Driever, W., Dowling,
606 J.E., 1995. A behavioral screen for isolating zebrafish mutants with visual system
607 defects. *Neurobiology* 92, 10545–10549.

608 Brustein, E., Saint-Amant, L., Buss, R.R., Chong, M., McDearmid, J.R., Drapeau, P., 2003. Steps
609 during the development of the zebrafish locomotor network. *J. Physiol.* 97, 77–86.

610 Cahill, D., Yuile, C., 1970. Tritium: some effects of continuous exposure in utero on
611 mammalian development. *Radiat Res* 44, 727–737. <https://doi.org/10.2307/3573152>

612 Carvan, M.J., Loucks, E., Weber, D.N., Williams, F.E., 2004. Ethanol effects on the developing
613 zebrafish: neurobehavior and skeletal morphogenesis. *Neurotoxicol. Teratol.* 26, 757–
614 768.

615 Chen, Y.-H., Huang, Y.-H., Wen, C.-C., Wang, Y.-H., Chen, W.-L., Chen, L.-C., Tsay, H.-J., 2008.
616 Movement disorder and neuromuscular change in zebrafish embryos after exposure to
617 caffeine. *Neurotoxicol. Teratol.* 30, 440–447.

618 Chhetri, J., Jacobson, G., Gueven, N., 2014. Zebrafish-on the move towards ophthalmological
619 research. *Eye* 28, 367–380. <https://doi.org/10.1038/eye.2014.19>

620 Dahm, R., Schonhaler, H.B., Soehn, A.S., van Marle, J., Vrensen, G.F.J.M., 2007.
621 Development and adult morphology of the eye lens in the zebrafish. *Exp. Eye Res.* 85,
622 74–89. <https://doi.org/10.1016/j.exer.2007.02.015>

623 Dhakal, S., Stevens, C., Sebbagh, M., Weiss, O., Frey, R., Adamson, S., Shelden, E., Inbal, A.,
624 Stenkamp, D., 2015. Abnormal retinal development in cloche mutant zebrafish. *Dev.*
625 *Dyn.* 244, 1439–1455. <https://doi.org/10.1002/dvdy.24322>

626 Drapeau, P., Saint-Amant, L., Buss, R.R., Chong, M., McDearmid, J.R., Brustein, E., 2002.
627 Development of the locomotor network in zebrafish. *Prog. Neurobiol.* 68, 85–111.

628 Ducros, L., Eyrolle, F., Della Vedova, C., Charmasson, S., Leblanc, M., Mayer, A., Babic, M.,
629 Antonelli, C., Mourier, D., Giner, F., 2018. Tritium in river waters from French
630 Mediterranean catchments: Background levels and variability. *Sci. Total Environ.* 612,
631 672–682.

632 Eaton, R.C., Farley, R.D., Kimmel, C.B., Schabtach, E., 1977. Functional development in the
633 Mauthner cell system of embryos and larvae of the zebra fish. *J. Neurobiol.* 8, 151–172.

634 Eyrolle-Boyer, F., 2017. Comportement et devenir du tritium dans l'environnement, in:
635 Actualisation Des Connaissances Acquisées Sur Le Tritium Dans l'environnement. pp. 22–
636 57.

637 Faucher, K., Floriani, M., Gilbin, R., Adam-Guillermin, C., 2012. Uranium-induced sensory
638 alterations in the zebrafish *Danio rerio*. *Aquat. Toxicol.* 124–125, 94–105.
639 <https://doi.org/10.1016/j.aquatox.2012.08.004>

640 Fraysse, B., Mons, R., Garric, J., 2006. Development of a zebrafish 4-day embryo-larval
641 bioassay to assess toxicity of chemicals. *Ecotoxicol. Environ. Saf.* 63, 253–267.

642 Gagnaire, B., Adam-Guillermin, C., Festarini, A., Cavalié, I., Della-Vedova, C., Shultz, C., Kim,
643 S.B., Ikert, H., Dubois, C., Walsh, S., Farrow, F., Beaton, D., Tan, E., Wen, K., Stuart, M.,
644 2017. Effects of in situ exposure to tritiated natural environments: A multi-biomarker
645 approach using the fathead minnow, *Pimephales promelas*. *Sci. Total Environ.* 599–600,
646 597–611. <https://doi.org/10.1016/j.scitotenv.2017.04.210>

647 Gagnaire, B., Cavalié, I., Pereira, S., Floriani, M., Dubourg, N., Camilleri, V., Adam-Guillermin,
648 C., 2015. External gamma irradiation-induced effects in early-life stages of zebrafish,
649 *Danio rerio*. *Aquat. Toxicol.* 169, 69–78. <https://doi.org/10.1016/j.aquatox.2015.10.005>

650 Giraudoux, P., Giraudoux, M.P., MASS, S., 2018. Package 'pgirmess.'

651 Goishi, K., Shimizu, A., Najarro, G., Watanabe, S., Rogers, R., Zon, L., Klagsbrun, M., 2006. A-
652 crystallin expression prevents -crystallin insolubility and cataract formation in the
653 zebrafish cloche mutant lens. *Dev. Dis.* 133, 2585–2593.
654 <https://doi.org/10.1242/dev.02424>

655 Gonen, R., German, U., Alfassi, Z.B., Priel, E., 2014. Production of DNA Double Strand Breaks
656 in Human Cells due to Acute Exposure to Tritiated Water (HTO). *Radiat. Prot.*

657 Hagger, J.A., Atienzar, F.A., Jha, A.N., 2005. Genotoxic, cytotoxic, developmental and survival
658 effects of tritiated water in the early life stages of the marine mollusc, *Mytilus edulis*.
659 *Aquat. Toxicol.* 74, 205–217. <https://doi.org/10.1016/j.aquatox.2005.05.013>

660 HPA, 2007. Review of Risks from Tritium. Report of the independent Advisory Group on
661 Ionising Radiation.

662 Huang, H., Huang, C., Wang, L., Ye, X., Bai, C., Simonich, M.T., Tanguay, R.L., Dong, Q., 2010.
663 Toxicity, uptake kinetics and behavior assessment in zebrafish embryos following
664 exposure to perfluorooctanesulphonic acid (PFOS). *Aquat. Toxicol.* 98, 139–147.
665 <https://doi.org/10.1038/jid.2014.371>

666 Hurem, S., Gomes, T., Brede, D.A., Lindbo Hansen, E., Mutoloki, S., Fernandez, C., Mothersill,
667 C., Salbu, B., Kassaye, Y.A., Olsen, A.K., Oughton, D., Aleström, P., Lyche, J.L., 2017.
668 Parental gamma irradiation induces reprotoxic effects accompanied by genomic
669 instability in zebrafish (*Danio rerio*) embryos. *Environ. Res.* 159, 564–578.
670 <https://doi.org/10.1016/j.envres.2017.07.053>

671 Hyodo-Taguchi, Y., Etoh, H., 1993. Vertebral Malformations in Medaka (Teleost Fish) after
672 Exposure to Tritiated Water in the Embryonic Stage. *Radiat. Res.* 135, 400–404.
673 <https://doi.org/10.2307/3578881>

674 ICRP, 2012. Protection of the Environment under Different Exposure Situations. ICRP
675 Publication 124. *Ann. ICRP* 43(1). <https://doi.org/10.1016/j.icrp.2006.06.001>

676 Jaeschke, B.C., Millward, G.E., Moody, A.J., Jha, A.N., 2011. Tissue-specific incorporation and
677 genotoxicity of different forms of tritium in the marine mussel, *Mytilus edulis*. *Environ.*
678 *Pollut.* 159, 274–280. <https://doi.org/10.1016/j.envpol.2010.08.033>

679 Jha, A.N., Dogra, Y., Turner, A., Millward, G.E., 2005. Impact of low doses of tritium on the
680 marine mussel, *Mytilus edulis*: Genotoxic effects and tissue-specific bioconcentration.
681 *Mutat. Res. - Genet. Toxicol. Environ. Mutagen.* 586, 47–57.
682 <https://doi.org/10.1016/j.mrgentox.2005.05.008>

683 Laskey, J., Parrish, J., Cahill, D., 1973. Some effects of lifetime Parental Exposure to Low
684 Levels of Tritium on the F2 Generation. *Radiat. Res.* 56, 171–179.
685 <https://doi.org/10.2307/3573801>

686 Lowry Dobson, R., Cooper, M., 1974. Tritium toxicity: Effect of low-level 3HOH exposure on
687 developing female germ cells in the mouse. *Radiat. Res.* 58, 91–100.

688 MacPhail, R.C., Brooks, J., Hunter, D.L., Padnos, B., Irons, T.D., Padilla, S., 2009. Locomotion
689 in larval zebrafish: Influence of time of day, lighting and ethanol. *Neurotoxicology* 30,
690 52–58.

691 Mathur-De Vre, R., Binet, J., 1984. Molecular aspects of tritiated water and natural water in
692 radiation biology. *Prog. Biophys. Mol. Biol.* 43, 161–193.

693 Mousseau, T.A., Møller, A.P., 2013. Elevated Frequency of Cataracts in Birds from Chernobyl.
694 *PLoS One* 8. <https://doi.org/10.1371/journal.pone.0066939>

695 OECD, 2013. Test No. 236: OECD Guidelines for the testing of chemicals. Fish
696 Embryo Acute Toxicity (FET) Test. <https://doi.org/10.1787/9789264203709-en>

697 Padilla, S., Hunter, D.L., Padnos, B., Frady, S., MacPhail, R.C., 2011. Assessing locomotor
698 activity in larval zebrafish: Influence of extrinsic and intrinsic variables. *Neurotoxicol.*
699 *Teratol.* 33, 624–630. <https://doi.org/10.1016/j.ntt.2011.08.005>

700 Parker, B., Connaughton, V.P., 2007. Effects of Nicotine on Growth And Development in
701 Larval Zebrafish. *Zebrafish* 4, 59–68. <https://doi.org/10.1089/zeb.2006.9994>

702 Pereira, S., Bourrachot, S., Cavalie, I., Plaire, D., Dutilleul, M., Gilbin, R., Adam-Guillermin, C.,
703 2011. Genotoxicity of acute and chronic gamma-irradiation on zebrafish cells and
704 consequences for embryo development. *Environ. Toxicol. Chem.* 30, 2831–2837.

705 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., Heisterkamp, S., Van Willigen, B., Maintainer,
706 R., 2017. Package ‘nlme.’ Linear Nonlinear Mix. Eff. Model. version 1–3.

707 R Core Team, 2017. R: A language and environment for statistical computing. R Foundation
708 for Statistical Computing, Vienna, Austria.

709 Rogakou, E.P., Pilch, D.R., Orr, A.H., Ivanova, V.S., Bonner, W.M., 1998. Double-stranded
710 Brekas Induce Histone H2AX phosphorylation on Serine 139. *J. Biol. Chem.* 273, 5858–
711 5868. <https://doi.org/10.1074/jbc.273.10.5858>

712 Saint-Amant, L., 2006. Development of Motor Networks in Zebrafish Embryos. *Zebrafish* 3,
713 173–190.

714 Scholz, S., Fischer, S., Gündel, U., Küster, E., Luckenbach, T., Voelker, D., 2008. The zebrafish
715 embryo model in environmental risk assessment—applications beyond acute toxicity
716 testing. *Environ. Sci. Pollut. Res.* 15, 394–404.

717 Selderslaghs, I.W.T., Hooyberghs, J., De Coen, W., Witters, H.E., 2010. Locomotor activity in
718 zebrafish embryos: A new method to assess developmental neurotoxicity. *Neurotoxicol.*
719 *Teratol.* 32, 460–471. <https://doi.org/10.1016/j.ntt.2010.03.002>

720 Spence, R., Gerlach, G., Lawrence, C., Smith, C., 2008. The behaviour and ecology of the
721 zebrafish, *Danio rerio*. *Biol. Rev.* 83, 13–34.

722 Stuart, M., Festarini, A., Schleicher, K., Tan, E., Kim, S.B., Wen, K., Gawlik, J., Ulsh, B., 2016.
723 Biological effects of tritium on fish cells in the concentration range of international
724 drinking water standards. *Int. J. Radiat. Biol.* 92, 563–571.
725 <https://doi.org/10.1080/09553002.2016.1222090>

726 Suyama, I., Etoh, H., Maruyama, T., 1981. Effects of ionizing radiation on the early
727 development of *Oryzias* eggs. *J. Radiat.* <https://doi.org/10.1269/jrr.22.125>

728 Svoboda, K.R., Vijayaraghavan, S., Tanguay, R.L., 2002. Nicotinic receptors mediate changes
729 in spinal motoneuron development and axonal pathfinding in embryonic zebrafish
730 exposed to nicotine. *J. Neurosci.* 22, 10731–10741.

731 Wang, X., Dong, Q., Chen, Y., Jiang, H., Xiao, Q., Wang, Y., Li, W., Bai, C., Huang, C., Yang, D.,
732 2013. Bisphenol A affects axonal growth, musculature and motor behavior in
733 developing zebrafish. *Aquat. Toxicol.* 142, 104–113.

734 Westerfield, M., 1995. *The zebrafish book: a guide for the laboratory use of zebrafish*

735 (Brachydanio rerio). University of Oregon press.

736

737 **List of figures**

738 **Figure 1:** Burst activity in control (white bars) and HTO-contaminated (grey bars) 24 hpf
739 embryos. The left bar plot represents control and contaminated 24 hpf embryos exposed at
740 0.4 mGy/h. Right bar plot represents control and HTO-contaminated 24 hpf embryos
741 exposed at 4 mGy/h (Permutation test pvalue: * < 0.05)

742 **Figure 2:** Heart-beats per second in control (white bars) and HTO-contaminated (grey bars)
743 organisms. 48 hpf larvae were exposed at 0.4 mGy/h (top left) and 4 mGy/h (bottom left). 72
744 hpf larvae were exposed at 0.4 mGy/h (top right) and 4 mGy/h (bottom right) (Permutation
745 test pvalue: * < 0.05)

746 **Figure 3:** Bar-plot of the comparison between the larvae basal velocity (mm/s) of control
747 (white bars) and HTO-contaminated (grey bars) 96 hpf larvae after exposure at (A) 0.4
748 mGy/h and (B) 4 mGy/h of HTO (Permutation test pvalue: * < 0.05)

749 **Figure 4:** Comparison between the velocity (mm/s) of 96 hpf larvae one second before
750 (white bars) and one second after (grey bars) the tactile stimulus. The left bar-plot
751 represents control and contaminated 96 hpf larvae exposed at 0.4 mGy/h. The right bar-plot
752 represents control and contaminated 96 hpf larvae exposed at 4 mGy/h (Permutation test
753 pvalue: * < 0.05)

754 **Figure 5:** Bar-plot representing the delta of the response to the tactile stimulus in control
755 (white bars) and contaminated (grey bars) in 96 hpf larvae. The left bar-plot represents
756 control and contaminated 96 hpf larvae exposed at 0.4 mGy/h. The right bar-plot represents
757 control and contaminated 96 hpf larvae exposed at 4 mGy/h (Permutation test pvalue: * <
758 0.05). Delta = after-before

759 **Figure 6:** Histological sections of control (left) and contaminated (right, 4 mGy/h) lens of 48
760 hpf, 72 hpf and 96 hpf larvae (from top to bottom). At 48 hpf (top), the lens is formed of a
761 monolayer of epithelial cells (dashed yellow lines) and primary lens fibers (PLF) with visible
762 nucleus. The cornea (C), the vitreous cavity (V) and blood vessels (BV) are also visible. At 72
763 hpf the lens fiber shows a more elongated pattern and at 96 hpf, the lens (L) show a more
764 homogeneous pattern. In all panels, ventral is down

765 **Figure 7:** Detection of gamma-H2AX foci per nucleus in 24 hpf embryos by whole mount-
766 immunostaining. Comparison of control organisms (white bars) and HTO-contaminated
767 organisms (grey bars). Embryos exposed at 0.4 mGy/h are represented in the left bar plot
768 and those exposed at 4 mGy/h in the right bar plot (Wilcoxon pvalue: * < 0.05)